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Is there a Case for Carbon-Based Border Tax Adjustment?

AN APPLIED GENERAL EQUILIBRIUM ANALYSIS

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ABSTRACT/RÉSUMÉ

Is there a case for carbon-based border tax adjustment? An applied general equilibrium analysis

Concern that unilateral greenhouse gas emission reductions could foster carbon leakage and undermine the international competitiveness of domestic industry has led to growing calls for carbon-based border-tax adjustments (BTAs). This paper uses a global general equilibrium model to assess the economic effects of BTAs and comes to three main conclusions. First, BTAs can reduce carbon leakage if the coalition of countries taking action to reduce emissions is small, because in this case leakage (while typically small) mainly occurs through international trade competitiveness losses rather than through declines in world fossil fuel prices that trigger rising carbon intensities outside the region taking action. Second, the welfare impacts of BTAs are small, and typically slightly negative at the world level. Third, and perhaps more strikingly, BTAs do not necessarily curb the output losses incurred by the domestic energy intensive-industries (EIIs) they are intended to protect in the first place. This is in part because taken as a whole, EIIs in industrialised countries make important use of carbon-intensive intermediate inputs produced by EIIs in other geographical areas. Another, deeper explanation is that EIIs are ultimately more adversely affected by carbon pricing itself, and the associated contraction in market size, than by any international competitiveness losses. These findings are shown to be robust to key model parameters, country coverage and design features of BTAs.

JEL classification: Q54; H25; D58

Keywords: Climate change; mitigation; border tax adjustment; computable general equilibrium model.

Y-a-t-il un argument en faveur d'une taxe carbone aux frontières ? Une analyse d'équilibre général

Les craintes que des réductions unilatérales d'émissions de gaz à effet de serre soient en partie compensées par des fuites de carbone tout en ayant un effet négatif sur la compétitivité des industries domestiques ont entraîné des appels croissants en faveur de taxes carbone aux frontières (TCFs). Cet article utilise un modèle d'équilibre général appliqué pour évaluer les effets économiques des TCFs et aboutit à trois conclusions. Premièrement, les TCFs peuvent réduire les fuites de carbone lorsque la coalition de pays prenant des mesures de réduction des émissions est réduite, car dans ce cas les fuites carbone (quoique typiquement faibles) se produisent essentiellement *via* des pertes de compétitivité internationale, plutôt que *via* des baisses du prix mondial des énergies fossiles qui entraînent une hausse de l'intensité en carbone dans le reste du monde. Deuxièmement, les impacts des TCFs sur le bien-être sont faibles, et typiquement légèrement négatifs au niveau mondial. Troisièmement, et peut-être de façon plus frappante, les TCFs n'atténuent pas nécessairement les pertes de production subies par les industries domestiques intensives en énergie (IIEs) qu'elles sont pourtant censées protéger. Cela tient en partie à ce que prises dans leur ensemble, les IIEs dans les pays industrialisés utilisent de façon importante des intrants intensifs en carbone produits par les IIEs d'autres zones géographiques. Une autre explication plus profonde est que les IIEs sont *in fine* davantage touchées par l'existence d'un prix du carbone lui-même, et par la contraction de la taille du marché qui s'en suit, que par de quelconques pertes de compétitivité internationale. Ces résultats s'avèrent robustes à des hypothèses alternatives concernant certains paramètres clé du modèle, les pays couverts et les modalités de mise en place des TCFs.

Classification JEL: Q54 ; H25 ; D58 *Mots-clé:* Changement climatique ; atténuation ; taxe carbone aux frontières ; modèle d'équilibre général calculable.

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TABLE OF CONTENTS

Abstract/Résumé.....	2
Is there a case for carbon-based border tax adjustment? An applied general equilibrium analysis	5
1. Introduction and executive summary	5
2. Two illustrative unilateral emission reduction scenarios	7
3. The impact of BTAs on carbon leakage, the output of EIIs and welfare	9
4. Generalisation of the results and sensitivity analysis.....	11
5. Conclusion	13
References	15
<i>Annex 1. An overview of the OECD ENV-linkages model.....</i>	<i>17</i>
References	20
<i>Annex 2. Constructing a business-as-usual baseline projection.....</i>	<i>21</i>
 Tables	
Table 1. Two benchmark scenarios of the impact of BTAs in 2030	8
Table 2. Impact of Border Tax Adjustments under alternative implementation assumptions	12
Table 3. Sensitivity analysis with respect to the values of key parameters	14
Table A1.1. ENV-Linkages model sectors.....	17
Table A1.2. ENV-Linkages model regions	18
 Figures	
Figure A2.1. Projected GHG emissions ¹ by country/region (2005-50, Gt CO ₂ eq).....	22

IS THERE A CASE FOR CARBON-BASED BORDER TAX ADJUSTMENT? AN APPLIED GENERAL EQUILIBRIUM ANALYSIS

By Jean-Marc Burniaux, Jean Chateau and Romain Duval¹

1. Introduction and executive summary

Anthropogenic climate change is a global public bad that calls for a global policy answer. Yet, partly reflecting strong free-riding incentives, the immediate prospects for a global carbon price addressing the negative externality associated with greenhouse gas (GHG) emissions are weak. Policy action is proceeding only gradually, with only some of the main emitting countries taking on binding policy measures. The COP15 held in Copenhagen in 2009 confirmed that global climate policy action will likely be built out of a collection of fragmented domestic commitments. At the same time, growing concern in industrialised countries that such unilateral reductions could foster “carbon leakage” and undermine the international competitiveness of domestic industries have led to growing calls for border-tax adjustments (BTAs) to “level the playing field”, particularly in the European Union (EU) and the United States (US). BTAs could take several forms, such as taxing imports or forcing importers to surrender emission allowances under domestic emission trading schemes (ETS). Computing the carbon content of imported goods that would be subject to BTAs could be undertaken in a number of ways. Regulators could attempt to estimate the direct carbon content of imports, the total embodied carbon including inputs in the value chain or the content of comparable domestic goods. None of these measures is straightforward.

Neither is the environmental rationale for BTAs, notwithstanding the participation incentives they might provide by punishing free-riding. There is a fairly broad consensus from the computable general equilibrium (CGE) literature that carbon leakage is limited, at least under plausible assumptions about key parameters such as carbon supply and trade elasticities (see in particular Burniaux and Oliveira Martins, 2000), and some papers have even explored the possibility of reverse leakage through endogenous technological change and international technology spillovers to non-mitigating countries (Grubb *et al.*, 2002; Gerlagh and Kuik, 2007).² There is a much clearer political economy rationale, however. Even

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1. The authors are Senior Economist at the OECD Environment Directorate, Economist at the OECD Environment Directorate and Head of Division at the OECD Economics Department, respectively. They would like to thank Cuauhtemoc Rebolledo-Gómez and Olivier Besson for excellent statistical and editing support, and Jan Corfee-Morlot, Rob Dellink, Jorgen Elmeskov, Stéphanie Jamet, Christine de la Maisonneuve, Helen Mountford, John Stephenson and Simon Upton for their input, suggestions and comments. Any errors are the responsibilities of the authors alone. The views expressed in this paper are those of the authors, and do not necessarily reflect those of the OECD or of its member countries.
 2. The focus of this literature, including the present paper, is on leakage across space. Some recent studies have stressed that more leakage might in fact occur across time through an inter-temporal channel. With finite fossil fuel endowments, pricing carbon today and announcing higher prices tomorrow gives carbon producers incentives to raise supply today, possibly leading to an increase rather than a decline in world emissions (Sinn, 2008). This mechanism is found to be even larger under incomplete geographical coverage of carbon pricing (Eichner and Pethig, 2009).

though leakage may be small overall, domestic energy-intensive industries exposed to international competition (EIIs) may still incur sizeable competitiveness and output losses from unilateral emission reduction action, especially in oligopolistic sectors producing an homogenous tradable good (Babikker, 2005; Demailly and Quirion, 2006, 2008a, 2008b). EIIs have been lobbying hard for BTAs in both the EU and the US, with the result that some recent BTA proposals focus essentially on these industries.

So far the economic debate on BTAs has mostly centered on administrative feasibility and consistency with WTO rules, and on compatibility with free trade more broadly (De Cendra, 2006; Demaret and Stewardson, 1994; Goh, 2004; Ismer and Neuhoﬀ, 2007; Perez, 2007; Stiglitz, 2006). However, surprisingly little economic analysis has been performed to assess the actual economic effects of BTAs. Based on earlier literature on the equivalence between origin and destination based sales taxes (Dosser, 1967; Krauss and Johnson, 1972; Shibata, 1967) and its implications for the effects of BTAs (Grossman, 1980; Whalley, 1979), Lockwood and Whalley (2008) discuss the economics of carbon-based BTAs. They stress that under a number of restrictive conditions (not least the absence of labour-leisure choice), and provided they apply similarly to all goods, (non carbon based) economy-wide BTAs have only nominal effects, without any real effects on production, consumption and trade. By contrast, carbon-based BTAs distort relative prices because the carbon content of goods differs and, therefore, they cannot be neutral. As a result, abstracting from their possible environmental effects, implementing a BTA along with a domestic carbon tax yields a welfare loss relative to a no-domestic-carbon-tax / no-BTA scenario. Whether BTAs also imply a welfare loss relative to a carbon-tax / no-BTA scenario is less straightforward in theory, as in this case BTAs may partly correct for the distortion associated with applying a carbon tax only to domestic goods.³ The effects of BTAs are therefore largely an empirical matter.

How BTAs are computed also matters. In principle, the domestic carbon price in the country acting unilaterally should be applied to the total (direct and indirect) carbon content of imported goods from a given country of origin. Although such calculation can in principle be completed within the framework of a world CGE model, it is unlikely to be feasible to implement it in practice. Partly for this reason, the handful of studies that have attempted to quantify the impact of carbon-based BTAs using CGE models have assumed different ways to calculate BTAs. For instance, Dong and Whalley (2009) build a small illustrative four-region two-sector computable general equilibrium (CGE) model, and find beneficial effects of BTAs on leakage but only small impacts on welfare and production for the EU and the US. This is partly because they assume BTAs to be applied based on the direct carbon content of domestically produced goods (in the EU and US), rather than on the (often much higher) carbon content of comparable imported goods. Mattoo *et al.* (2009) explore a broader range of options using a more detailed CGE framework, and find larger impacts of BTAs in some scenarios. But here again, it is a matter of how BTAs are modelled. In the benchmark scenarios of the present paper, the carbon content used to calculate BTAs is that of the country of origin of the imported good, and it includes both the direct carbon content based on fossil fuel inputs and the part of the indirect content that corresponds to fossil fuel inputs into the electricity used as intermediate consumption in the particular sector and country subject to the BTA. Alternative set-ups are also considered as part of an extensive sensitivity analysis.

This paper uses a global recursive-dynamic CGE model, ENV-Linkages, to assess the impacts of BTAs on leakage, competitiveness and welfare. Given the empirical nature of the issue, and the central role played by interactions across countries and sectors through trade and fossil fuel price channels, an applied global CGE framework appears to be the appropriate analytical tool. Two benchmark scenarios are

3. This is over and above any potential environmental benefits – resulting from reduced carbon leakage and thereby increased global emission reductions – from BTAs, which are not included in the welfare analysis performed in this paper. However, such gains are unlikely to radically alter our conclusions, as the unilateral targets considered below amount to a modest mitigation of worldwide emissions, a small share of which is subject to leakage.

considered under which either the EU alone or the group of Annex I countries under the Kyoto Protocol (mostly industrialised countries) cut their emissions by 20% by 2020 and 50% by 2050 relative to 2005 levels. Consistent with recent legislative initiatives and political economy fundamentals, special attention is paid to EIIs throughout the analysis. Importantly, a broad-based sensitivity analysis is performed to assess the robustness of the main results to key model parameters, targets, countries and design features of BTAs.

Two main conclusions stand out. First, BTAs can reduce carbon leakage for small coalitions of acting countries such as the EU, because in this case leakage (while typically small) mainly occurs through international trade competitiveness losses rather than through declines in world fossil fuel prices. However, the need for, and the effectiveness of BTAs declines rapidly with the size of the coalition as BTAs address a smaller share of an ever smaller rate of leakage. Second, the economic effects of BTAs are small. They have negligible welfare effects both worldwide and for countries that impose them. This is not wholly unexpected given that their effects are theoretically ambiguous. Perhaps more strikingly, BTAs do not necessarily curb the output losses incurred by the domestic EIIs they are intended to support in the first place. This is in part because in industrialised countries, EIIs make important use of carbon-intensive intermediate inputs produced by EIIs in other geographical areas. Another, deeper explanation is that EIIs are ultimately more adversely affected by carbon pricing itself – which is needed to achieve a cost-effective reduction in emissions – and the associated contraction in market size, than by any international competitiveness losses and the associated reduction in market share.

The rest of this paper proceeds as follows. Section 2 assesses leakage, welfare and competitiveness losses from unilateral emission reduction action under the two benchmark scenarios. Section 3 explores the effects of introducing BTAs under these scenarios, and Section 4 performs sensitivity analysis of the main results to key model parameters, targets and design features of BTAs. Section 5 concludes.

2. Two illustrative unilateral emission reduction scenarios

The assessment of the economic effects of BTAs relies in a first step – before carrying out extensive sensitivity analysis in a second step – on two benchmark climate policy scenarios, under which either the EU alone or Annex I countries under the Kyoto Protocol as a whole cut their emissions by 20% by 2020 and by 50% by 2050, relative to 2005 levels. At the 2020 horizon, the former scenario (scenario “EU no BTA” in Table 1) corresponds in fact to the official EU emission reduction target.⁴ The latter scenario (scenario “A1 no BTA” in Table 1) is more illustrative, and aims at exploring possible differences in the magnitude of leakage and the economic effects of BTAs between smaller and larger coalitions of acting countries. Reflecting the likely magnitude and unpredictability of long-term changes in the structure of the world economy, as well as the very low probability that a small number of countries will act alone over a horizon of several decades anyway, the main focus is on the 2030 horizon. Results at the 2020 horizon are qualitatively similar but quantitatively smaller, reflecting a less stringent target – and therefore a lower carbon price – at this horizon.

4. The EU has indicated that this target could be raised from 20% to 30% if other countries took on “comparable efforts”. Note that the EU target is defined as a percentage of 1990 levels, but since emissions in 2005 are globally identical to 1990 for the region as a whole, this difference does not matter.

Table 1. Two benchmark scenarios of the impact of BTAs in 2030
(Emissions reduction profile: -20% in 2020 and -50% in 2050 for acting countries, relative to 2005 levels)

Policy scenario	Carbon tax (USD/t CO ₂)	Leakage rate (%)	Equivalent variation in income			EII output			World GHG emissions	
			World	non-acting countries	acting countries	World	non-acting countries	acting countries		
% change in 2030 with respect to the baseline										
Reference scenarios										
EU no BTA	EU countries acting alone	61	7.9	-0.4	-0.1	-1.4	-0.2	0.2	-2.2	-2.4
EU BTA dir	EU countries acting alone + BTA on direct carbon contents	63	1.0	-0.4	-0.1	-1.2	-0.3	0.2	-2.5	-2.6
EU BTA ind	EU countries acting alone + BTA on direct and indirect carbon contents	63	-1.4	-0.4	-0.2	-1.1	-0.3	0.1	-2.5	-2.6
A1 no BTA	AnnexI countries acting alone	74	5.9	-1.2	-0.5	-1.6	-0.9	1.1	-3.2	-12.6
A1 BTA dir	AnnexI countries acting alone + BTA on direct carbon contents	74	3.4	-1.2	-0.8	-1.5	-1.1	0.9	-3.4	-13.0
A1 BTA ind	AnnexI countries acting alone + BTA on direct and indirect carbon contents	74	2.2	-1.3	-1.0	-1.4	-1.2	0.6	-3.3	-13.1

Source: OECD ENV-Linkages model

Scenarios EU no BTA and A1 no BTA are both simulated using the OECD's ENV-Linkages model, a global recursive-dynamic CGE model featuring 12 world regions and 22 sectors, and including both CO₂ and non-CO₂ GHGs. The main features of the model are discussed in Annex 1 and in greater detail in Burniaux and Chateau (2008), while the baseline (no-carbon-price / no-BTA) scenario that underpins it, is briefly described in Annex 2 and in full in Duval and de la Maisonnette (2010). The welfare impacts of policy action are measured relative to baseline using the Hicksian equivalent variation in income to assess changes in real income. These utility-based welfare measures do not incorporate the impacts of climate change, which are not covered in ENV-Linkages. While such impacts are subject to broad uncertainty and are small anyway at the 2030 horizon of this paper, this should be borne in mind when interpreting the results from the welfare analyses performed below. Throughout the analysis special attention is also paid to EIIs, which include here chemicals, non-ferrous metals, fabricated metal products, iron and steel, pulp and paper and non-metallic mineral products. The effects of both scenarios on leakage, real income (Hicksian equivalent variation in income) and the real output of EIIs are presented in Table 1. Leakage is found to be small and to decline with coalition size, reaching just 8% of the decline in EU emissions by 2030 in scenario "EU no BTA", and 6% of the decline in Annex I emissions in scenario "A1 no BTA".⁵ Moreover, it is not only the magnitude but also the nature of leakage that changes with the size of the coalition. Indeed leakage can arise through two main channels: the international trade channel, as carbon-intensive industries in acting countries lose market shares to their foreign competitors and/or relocate capital in non-acting countries; the fossil fuel price channel, as emission reduction efforts in acting countries lower world demand for fossil fuels, thereby inducing a price decline that triggers greater fossil fuel use and higher GHG emissions in non-participating countries. The wider the country coverage, the smaller the market share losses affecting EIIs in participating countries (the international trade channel of leakage), but the larger the impact of policy action on international fossil fuel prices (the fossil fuel price channel of leakage).

The negative welfare impact of achieving the target is smaller for the EU (scenario "EU no BTA") than for Annex I countries as a whole (scenario "A1 no BTA"), reflecting the faster baseline emissions growth and higher carbon intensity of the latter group of countries. As would be expected, qualitatively similar, but quantitatively larger impacts are found for the output of EIIs.

3. The impact of BTAs on carbon leakage, the output of EIIs and welfare

The impacts of implementing a BTA under the two benchmark scenarios are now simulated using the ENV-Linkages model. BTAs should in principle apply to the actual carbon content of imported goods, rather than to the carbon content of comparable domestic goods. Therefore the former set up is retained, although the latter is also analysed in Section 4 as it might be easier to implement in practice. Another issue is the extent to which BTAs would apply not only to the direct carbon content of goods but also to their indirect content, *i.e.* to the carbon content of the inputs used to produce these goods. While in theory they could be simulated with ENV-Linkages as they are in Mattoo *et al.* (2009), in practice accurately calculating the full indirect carbon content of goods is likely to be impossible, especially given the length and complexity of value added chains in an increasingly globalised production process. Therefore two more realistic alternatives are considered here, namely BTAs applied either only to the direct carbon content of (imported) goods (Scenarios "EU BTA dir" and "A1 BTA dir") or to the direct plus part of the indirect content *via* the carbon content of electricity inputs only (Scenarios "EU BTA ind" and "A1 BTA ind").

The results are presented in Table 1. BTAs appear to reduce the carbon leakage from the unilateral emission reduction measures of small coalitions, but their effectiveness declines rapidly with coalition size.

5. Note that the lower leakage rate in scenario "A1 no BTA" compared with scenario "EU no BTA" is not entirely straightforward *a priori*, because the 20% emission reduction objective under both scenarios implies a larger absolute cut (in giga tons CO₂ equivalent, Gt CO₂-eq) than in the former.

Indeed the (limited) leakage problem is fully addressed in the EU case (scenarios “EU BTA dir” and “EU BTA ind”),⁶ but less so when a larger coalition such as Annex I countries takes action (scenarios “A1 BTA dir” and “A1 BTA ind”). This is primarily because under smaller coalitions, leakage arises comparatively more from international competitiveness losses than through a decline in world fossil fuel prices, making BTAs a more effective tool since they address the former but not the latter channel. As a result of the leakage reduction, the environmental effectiveness of unilateral actions in terms of world emissions reduction is improved, although only marginally.

By contrast, the economic effects of BTAs are found to be fairly small overall in both scenarios. BTAs mitigate the welfare losses from carbon pricing for acting countries, but at the world level these gains are roughly offset by additional losses in the rest of the world. When BTAs are applied at the level of Annex I countries, the real income loss incurred by non-Annex I countries is somewhat higher, reaching 1% of their real income. As noted above, negligible welfare effects from BTAs in the countries that implement them are not entirely unexpected from theory, as their impact is *a priori* ambiguous and driven by the difference in – not the level of – the tariffs applied across goods.⁷

Perhaps more surprisingly, despite some effectiveness in reducing leakage, BTAs are not found to curb the output losses of EIIIs taken as a whole (relative to a carbon tax/no BTA scenario). The output losses incurred by EIIIs in both scenarios are roughly unchanged when either the direct or both the direct and indirect carbon contents of imports are subject to a border tariff. This is in part because several factors contribute to offset the positive output effects of the market share gains associated with BTAs: first and foremost, since domestic EIIIs (taken as a whole) in industrialised countries rely heavily on imported inputs produced by EIIIs at a different level of the value added chain in other countries, BTAs increase the production costs of domestic EIIIs;⁸ second, realistic but incomplete forms of BTAs such as those considered here, which do not cover the full indirect carbon content of imports, do not fully address the competitiveness losses of domestic EIIIs; third, the presence of BTAs induces a slight increase in the carbon price to meet the domestic emission target, which further increases the production costs of EIIIs. However, as will become apparent in Section 4 below, the single most important factor behind the lack of effectiveness of BTAs to support domestic EIIIs is that these industries are ultimately more adversely affected by the existence of carbon pricing itself than by any international competitiveness losses.

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6. The EU is even found to experience “negative leakage” once BTAs are implemented. This is because the supply of coal is more elastic than that of crude oil in the benchmark calibration of ENV-Linkages, making coal relatively more expensive in world international markets when emission reduction measures – both direct and indirect such as BTAs – are taken by a reasonably large area such as the EU. This induces a substitution away from more carbon-intensive coal in non-participating countries, and therefore a decline in their emissions that amounts to negative carbon leakage.
 7. In the literature, the impact of GHG mitigation policies is sometimes reported in terms of GDP changes, although GDP is a less appropriate metric to measure welfare changes than the real income indicator used in Table 1. When expressed in terms of GDP, the impact of BTAs implies a slight *increase* in economic costs for acting countries, for instance from a GDP loss of 0.8% to 0.9% in 2030 in the scenario where the EU is acting alone. Real-exchange-rate and international trade changes account for this difference between the GDP and real income effects of BTAs. However, both indicators convey the same message that the economic effects of BTAs are small in the countries that implement them.
 8. Partly for this reason, unlike other EIIIs, the mineral (including cement) and non-ferrous metal industries – which rely less than other EIIIs on carbon-intensive imported inputs – are found to gain somewhat from BTAs.

4. Generalisation of the results and sensitivity analysis

In order to assess the extent by which the results from Section 3 can be generalized, sensitivity analysis is carried out by applying BTAs to other OECD countries, exploring alternative options for implementing them and considering alternative values for some critical model parameters.

The first panel of Table 2 reports impacts of BTAs under unilateral action by the USA (scenarios “US no BTA” and “US BTA ind”) and Japan (scenarios “JPN no BTA” and “JPN BTA ind”). While US BTAs yield a significantly smaller reduction in carbon leakage compared with the EU BTAs⁹ – from 12 to 9% in 2030, other results regarding real income and EII output effects are in line with those from the benchmark scenarios of Section 3. In particular, BTAs have little impact on mitigation costs and are ineffective in curbing the output losses incurred by EII, for reasons discussed above.

The second panel of Table 2 examines the consequences of alternative ways of implementing BTAs. For instance BTAs could be based on the carbon content of domestic rather than imported goods, as this would likely improve the practical feasibility of BTAs. Using domestic carbon contents (scenarios “EU BTA ind dom” and “A1 BTA ind dom” in Table 2) implies for the EU and Annex I countries a much smaller reduction in carbon leakage compared with the two benchmark scenarios of Section 3. In both cases, domestic carbon contents in acting countries are lower than those of imported goods due to a more efficient use of fossil fuels, and this results into smaller BTAs. While the environmental effectiveness of applying BTAs on this basis is therefore reduced, the economic costs of action in terms of real income and EII output losses are unchanged or even slightly increased.

BTAs on imports alone only partly address competitiveness concerns in acting countries as they fail to compensate for the competitiveness losses that domestic EII would incur on their export markets. In principle, this issue could be addressed through a symmetrical treatment of EII imports and exports, which would involve BTAs on imports along with some exemption of EII exports from the domestic carbon price. In practice, however, such a measure is found to have little impact on EII output losses (scenarios “EU BTA ind sub” and “A1 BTA ind sub” in Table 2), while increasing the overall economic losses incurred by acting countries, due to the higher carbon price that results from the narrower coverage of carbon pricing. As carbon leakage tends to increase with a higher carbon price, the environmental effectiveness of BTAs is also slightly reduced.

Similarly, restricting mitigation action to CO₂ emissions from fuel combustion only (rather than comprehensive mitigation across all greenhouse gases) requires much higher taxes (scenarios “EU no BTA CO₂” and “A1 no BTA CO₂” in Table 2) and implies significantly higher leakage rates: leakage rates calculated for all GHGs in 2030 reach 14% for the EU and 9% for Annex I countries, versus 8% and 6% respectively if mitigation involves all GHGs. This illustrates the role of CO₂ and the importance of the world energy markets in generating carbon leakages. The impact of BTAs (scenarios “EU BTA ind CO₂” and “A1 BTA ind CO₂”) under CO₂ mitigation only is in line with previous results, namely: the leakage rate is reduced (especially in the case of the EU); the real income loss in acting countries is slightly lower and offset at the world level by higher losses in other countries; there is no reduction in the output losses incurred by EII in acting countries.

9. This reflects the smaller weight of the competitiveness versus fossil fuel component in the leakage generated by US action, as the US economy is both less open to trade and accounts for a bigger share of world fossil fuel consumption than the EU.

Table 2. Impact of Border Tax Adjustments under alternative implementation assumptions

Policy scenario		Carbon tax (USD/t CO ₂)	Leakage rate (%)	Equivalent variation in income			EII output			World GHG emissions
				World	non-acting countries	acting countries	World	non-acting countries	acting countries	
% change in 2030 with respect to the baseline										
Alternative countries										
US no BTA	USA acting alone	73	11.8	-0.4	-0.1	-1.2	-0.3	0.6	-4.6	-5.2
US BTA ind	USA acting alone + BTA	75	8.6	-0.4	-0.2	-1.0	-0.4	0.4	-4.6	-5.4
JPN no BTA	Japan acting alone	30	12.5	-0.1	0.0	-0.4	0.0	0.1	-1.4	-0.5
JPN BTA ind	Japan acting alone + BTA	31	5.5	-0.1	0.0	-0.3	0.0	0.1	-1.4	-0.6
Alternative options for implementing BTAs										
EU BTA ind dom	EU acting alone + BTA based on domestic carbon content	62	5.2	-0.4	-0.1	-1.3	-0.3	0.2	-2.5	-2.5
A1 BTA ind dom	Annex I countries acting alone + BTA based on domestic carbon content	74	4.5	-1.3	-0.7	-1.6	-1.1	1.1	-3.6	-12.8
EU BTA ind sub	EU acting alone + BTA exempting exports	64	-0.9	-0.4	-0.1	-1.3	-0.3	0.2	-2.3	-2.6
A1 BTA ind sub	AnnexI countries acting alone + BTA exempting exports	75	3.6	-1.3	-0.7	-1.6	-1.2	0.8	-3.4	-12.9
EU no BTA CO2	EU acting alone with CO ₂ only	79	13.5	-0.4	-0.1	-1.7	-0.2	0.2	-2.1	-1.7
EU BTA ind CO2	EU acting alone with CO ₂ only + BTA	82	0.9	-0.5	-0.2	-1.4	-0.3	0.2	-2.5	-1.9
A1 no BTA CO2	AnnexI countries acting alone with CO ₂ only	86	9.1	-1.2	-0.5	-1.6	-0.9	1.3	-3.3	-9.8
A1 BTA ind CO2	AnnexI countries acting alone with CO ₂ only + BTA	87	4.7	-1.3	-1.0	-1.5	-1.2	0.9	-3.4	-10.4
EU BTA ind diag	EU acting alone with BTA and exempting diagonal imported input	64	-1.6	-0.4	-0.2	-1.1	-0.3	0.1	-1.9	-2.6
A1 BTA ind diag	AnnexI countries acting alone with BTA and exempting diagonal imported input	74	2.2	-1.3	-1.0	-1.4	-1.1	0.5	-3.0	-13.1

Source: OECD ENV-Linkages model

The last two scenarios (scenarios “EU BTA ind diag” and “A1 BTA ind diag”) in Table 2 show that a marginally more effective way to curb EII’s output losses in acting countries would be to exempt from the BTA their imported inputs of EII products originating from trading partners. However, such an approach would reduce EII’s output losses only marginally compared with the benchmark scenarios, from 2.2% to 1.9% in 2030 in the EU and from 3.3% to 3% in Annex I countries. The small magnitude of this gain confirms that the primary driver of EII’s output losses in countries that take mitigation action is the impact of carbon pricing on production costs, rather than any loss in international competitiveness.

Finally, Table 3 shows that these results are reasonably robust to alternative values of key parameters. In a nutshell, alternative parameter values do alter the environmental and economic effects of carbon taxes, but they have a more limited influence on the impact of BTAs, which is measured relative to those carbon price scenarios. A more (less) elastic supply of fossil fuels at the world level implies lower (higher) leakage from unilateral mitigation action. The effectiveness of BTAs in reducing carbon leakage is reduced under a less elastic fossil fuel supply, especially in the scenario where Annex I countries act alone (scenarios “A1 no BTA” and “A1 BTA ind” in Table 3). However, previous results regarding the economic effects of BTAs still hold. The degree of product differentiation in world trade markets also influences the amount of carbon leakage. If products from different origins are more substitutable (as simulated by raising the values of the Armington elasticities), carbon leakage is higher and the loss of competitiveness incurred by EII’s results in a larger output loss (scenarios “EU no BTA” in Table 3 and Table 1). Still, the environmental and economic impacts of the BTAs remain roughly unchanged.

5. Conclusion

As industrialised countries implement or consider more stringent unilateral constraints on domestic GHG emissions, the political momentum for BTAs to address carbon leakage and “level the playing field” between their EII’s and their unconstrained foreign competitors might grow. A small body of recent economic research that builds on earlier literature on border adjustments points to ambiguous and probably small welfare effects of BTAs a priori, but this has yet to be backed by fully-fledged applied analyses. This paper began to fill this gap by using a global recursive-dynamic CGE model, ENV-Linkages, to assess the potential impacts of BTAs on leakage, competitiveness and welfare. Illustrative unilateral emission reduction scenarios with and without BTAs are explored, and extensive sensitivity analysis is performed in order to assess the robustness of the results to targets, countries, design features of BTAs and key parameters such as fossil fuel supply or international trade elasticities. A robust finding across all simulations is that BTAs have only small welfare effects. They have also typically no beneficial impact on the output of the EII’s they are intended to support in the first place. BTAs primarily reduce the demand for, and thereby the output of the foreign competitors of domestic EII’s, leading to a mechanical increase in the global market share of domestic EII’s. However this does not bring any output gains since the positive impact of competitiveness gains is typically offset by a rise in production costs, and both effects are small anyway compared with the output losses associated with the existence of a carbon price. Under most circumstances, BTAs are more effective at reducing carbon leakage, and the environmental gains from lower global emissions are not factored into the welfare analysis performed in this paper. However, such gains are unlikely to radically alter the conclusions, as the unilateral targets considered here amount to a modest mitigation of worldwide emissions, a small share of which is subject to leakage. Finally, it should be stressed that the findings from this paper do not incorporate, and therefore come over and above existing concerns regarding the administrative costs and retaliation risks that BTAs may entail.

Table 3. Sensitivity analysis with respect to the values of key parameters

Policy scenario		Carbon tax (USD/t CO ₂) in 2030	Leakage rate (%) in 2030	% change in 2030 with respect to the baseline						
				Equivalent variation in income			EII output			World GHG emissions
				World	non-acting countries	acting countries	World	non-acting countries	acting countries	
Sensitivity to price elasticity of fossil fuel supply										
EU no BTA	Fossil fuel supply more elastic	59	5.2	-0.4	-0.1	-1.5	-0.2	0.2	-2.2	-2.5
EU BTA ind	Fossil fuel supply more elastic	61	-4.1	-0.4	-0.2	-1.2	-0.4	0.1	-2.4	-2.7
EU no BTA	Fossil fuel supply less elastic	66	13.6	-0.3	-0.1	-1.4	-0.2	0.3	-2.3	-2.2
EU BTA ind	Fossil fuel supply less elastic	68	4.3	-0.4	-0.2	-1.1	-0.3	0.2	-2.6	-2.5
A1 no BTA	Fossil fuel supply more elastic	70	4.2	-1.3	-0.5	-1.7	-1.0	1.0	-3.2	-12.9
A1 BTA ind	Fossil fuel supply more elastic	70	2.8	-1.3	-0.7	-1.7	-1.2	0.9	-3.6	-13.0
A1 no BTA	Fossil fuel supply less elastic	80	9.5	-1.1	-0.4	-1.4	-0.9	1.3	-3.3	-12.1
A1 BTA ind	Fossil fuel supply less elastic	81	8.1	-1.1	-0.6	-1.4	-1.1	1.3	-3.7	-12.3
Sensitivity to Armington Elasticities (AE)										
EU no BTA	Low AE for manufacturing goods in all countries	62	7.5	-0.4	-0.1	-1.4	-0.2	0.2	-2.0	-2.4
EU BTA ind	Low AE for manufacturing goods in all countries	64	-2.1	-0.4	-0.2	-1.1	-0.4	0.1	-2.4	-2.6
EU no BTA	High AE for manufacturing goods in all countries	61	8.8	-0.4	-0.1	-1.5	-0.3	0.3	-2.5	-2.4
EU BTA ind	High AE for manufacturing goods in all countries	63	-0.4	-0.4	-0.2	-1.2	-0.4	0.2	-2.6	-2.6
A1 no BTA	Low AE for manufacturing goods in all countries	74	5.3	-1.2	-0.6	-1.5	-1.0	0.8	-3.0	-12.7
A1 BTA ind	Low AE for manufacturing goods in all countries	74	1.5	-1.3	-1.1	-1.4	-1.3	0.4	-3.3	-13.2
A1 no BTA	High AE for manufacturing goods in all countries	73	6.9	-1.2	-0.4	-1.6	-0.9	1.6	-3.7	-12.5
A1 BTA ind	High AE for manufacturing goods in all countries	73	3.2	-1.3	-0.9	-1.5	-1.2	0.9	-3.5	-13.0

Source: OECD ENV-Linkages model (spring 2010 baseline)

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ANNEX 1. AN OVERVIEW OF THE OECD ENV-LINKAGES MODEL

The OECD ENV-Linkages model is a recursive dynamic neo-classical general equilibrium model. It is the successor to the OECD GREEN model for environmental studies (Burniaux, *et al.* 1992; Burniaux, 2000). The model is documented in Burniaux and Chateau (2008). Previous works using ENV-Linkages extensively include two books: OECD (2008) and OECD (2009).

ENV-Linkages is a global economic model built primarily on a database of national economies. In the version of the model used here, the world economy is divided in 12 countries/regions, each with 25 economic sectors (Tables A1.1 and A1.2), including five different technologies to produce electricity. The core of the static equilibrium is formed by the set of Social Account Matrices (SAMs) that describes how economic sectors are linked; these are based on the GTAP database (currently using version 6.2). A fuller description of the database can be found at Dimaranan (2006). Many key parameters are set on the basis of information drawn from various empirical studies and data sources (see Burniaux and Chateau, 2008).

Table A1.1. ENV-Linkages model sectors

1) Rice	14) Food Products
2) Other crops	15) Other Mining
3) Livestock	16) Non-ferrous metals
4) Forestry	17) Iron & steel
5) Fisheries	18) Chemicals
6) Crude Oil	19) Fabricated Metal Products
7) Gas extraction and distribution	20) Paper & Paper Products
8) Fossil Fuel Based Electricity	21) Non-Metallic Minerals
9) Hydro and Geothermal electricity	22) Other Manufacturing
10) Nuclear Power	23) Transport services
11) Solar& Wind electricity	24) Services
12) Renewable combustibles and waste electricity	25) Construction & Dwellings
13) Petroleum & coal products	26) Coal

All production in ENV-Linkages is assumed to operate under cost minimisation with an assumption of perfect markets and constant returns to scale technology. The production technology is specified as nested Constant Elasticity of Substitution (CES) production functions in a branching hierarchy. Each sector uses intermediate inputs – including energy inputs - and primary factors (labour, capital, land and natural resources). For each good or service, output is produced by different production streams which are differentiated by capital vintage (old and new). The substitution possibilities among production factors are assumed to be higher with the *new* than with the *old* capital vintages – technology has a putty/semi-putty specification. Capital accumulation is modelled according to the traditional Solow/Swan neo-classical growth model.

Table A1.2. ENV-Linkages model regions

ENV-Linkages regions	GTAP countries/regions
1) Australia & New Zealand	Australia, New Zealand
2) Japan	Japan
3) Canada	Canada
4) United States	United States
5) European Union 27 & EFTA	Austria, Belgium, Denmark, Finland, Greece, Ireland, Luxembourg, Netherlands, Portugal, Sweden, France, Germany, United Kingdom, Italy, Spain, Switzerland, Rest of EFTA, Czech Republic, Slovakia, Hungary, Poland, Romania, Bulgaria, Cyprus, Malta, Slovenia, Estonia, Latvia, Lithuania
6) Brazil	Brazil
7) China	China, Hong Kong
8) India	India
9) Russia	Russian Federation
10) Oil-exporting countries	Indonesia, Venezuela, Rest of Middle East, Islamic Republic of Iran, Rest of North Africa, Nigeria
11) Non-EU Eastern European countries	Croatia, Rest of Former Soviet Union
12) Rest of the world	Korea, Taiwan, Malaysia, Philippines, Singapore, Thailand, Viet Nam, Rest of East Asia, Rest of Southeast Asia, Cambodia, Rest of Oceania, Bangladesh, Sri Lanka, Rest of South Asia, Pakistan, Mexico, Rest of North America, Central America, Rest of Free Trade Area of Americas, Rest of the Caribbean, Colombia, Peru, Bolivia, Ecuador, Argentina, Chile, Uruguay, Rest of South America, Paraguay, Turkey, Rest of Europe, Albania, Morocco, Tunisia, Egypt, Botswana, Rest of South African Customs Union, Malawi, Mozambique, Tanzania, Zambia, Zimbabwe, Rest of Southern African Development Community, Mauritius, Madagascar, Uganda, Rest of Sub-Saharan Africa, Senegal, South Africa.

The energy bundle is of particular interest for analysis of climate change issues. Energy is a composite of fossil fuels and electricity. In turn, fossil fuel is a composite of coal and a bundle of the “other fossil fuels”. At the lowest nest, the composite “other fossil fuels” commodity consists of crude oil, refined oil products and natural gas. The value of the substitution elasticities are chosen as to imply a higher degree of substitution among the other fuels than with electricity and coal.

World trade is based on a set of regional bilateral flows. Allocation of trade between partners responds to changes in relative prices between regions. The basic assumption is that imports originating from different regions are imperfect substitutes (Armington specification). Each region runs a fixed current-account surplus (or deficit).

The ENV-Linkages model has a simple recursive-dynamic structure, where households base their decisions on static expectations concerning prices and quantities. Household consumption demand and savings are implemented through an “Extended Linear Expenditure System”. Since consumers are not represented with forward-looking behavior, some care needs to be exercised in studying policies that consumers may reasonably be expected to anticipate – either the policy itself or its consequences. In each period, investment net-of-economic depreciation is equal to the sum of government savings, consumer savings and net capital flows from abroad.

The government in each region collects various kinds of taxes in order to finance government expenditures. Aggregate government expenditures are linked to real GDP. Assuming fixed public savings (or

deficits), the government budget is balanced through the adjustment of the income tax on consumer income.

CO₂ emissions from combustion of energy are directly linked to the use of different fuels in production. Other GHG emissions are linked to output in a way similar to Hyman *et al.* (2002). The following non-CO₂ emission sources are considered: *i*) methane from rice cultivation, livestock production (enteric fermentation and manure management), coal mining, crude oil extraction, natural gas and services (landfills); *ii*) nitrous oxide from crops (nitrogenous fertilizers), livestock (manure management), chemicals (non-combustion industrial processes) and services (landfills); *iii*) industrial gases (SF₆, PFC's and HFC's) from chemicals industry (foams, adipic acid, solvents), aluminum, magnesium and semi-conductors production.

For studying the impacts of climate change policy, four types of instruments have been developed: 1) GHG taxes, global or specific by sectors, gases or emission sources; 2) tradable emission permits (with flexibility between regions and sectors); 3) offsets (including a stylised version of the Clean Development Mechanism); and 4) regulatory policy. Taxes and tradable permits are applied directly to GHG emissions. Offsets are driven by an exogenous limit on demand for offset credits and competition between potential suppliers. Regulatory policy has been introduced in the model through quantity constraints (Burniaux, *et al.* 2008).

Market goods equilibria imply that, on the one side, the total production of any good or service is equal to the demand addressed to domestic producers plus exports; and, on the other side, the total demand is allocated between the demands (both final and intermediary) addressed to domestic producers and the import demand. The general equilibrium framework ensures that a unique set of relative prices emerges such that demand equals supply in all markets simultaneously (*i.e.* across all regions, commodities, and factors of production). All prices are expressed relatively to the numéraire of the price system that is chosen as the index of OECD manufacturing exports prices. Implementation of a policy in the model leads to a new equilibration process and thus a new set of equilibrium prices and quantities to compare with the original equilibrium.

The process of calibration of the ENV-Linkages model is broken down into three stages (cf. Burniaux and Chateau, 2008). First, a number of parameters are calibrated, given some elasticity values, on base-year (2001) values of variables. Second, the 2001 database is updated to 2005 by simulating the model dynamically to match historical trends over the period 2001-05; thus all variables are expressed in 2005 real USD. Third, the baseline projection until 2050 is based on convergence assumptions about labour productivity and other socio-economic drivers (demographic trends, future trends in energy prices and energy efficiency gains), as further described in Duval and de la Maisonneuve (2010). The baseline has been adjusted to incorporate the effects of the economic crisis of 2008-09. In addition, the baseline assumes no new climate policies, but does include other government policies for instance on energy policy as included in the energy projections of the IEA (2009).¹⁰ It thus provides a benchmark against which policy scenarios aimed at achieving emission cuts can be assessed.

10. The baseline simulation also contains the assumption that the EU Emission Trading System is implemented over the period 2006-12, assuming a permits price that will rise gradually from 5 to 25 constant \$US in 2012.

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ANNEX 2. CONSTRUCTING A BUSINESS-AS-USUAL BASELINE PROJECTION

Assumptions about drivers of GDP

Baseline economic scenarios underlying climate change projections – such as those developed for the IPCC (Nakicenovic *et al.*, 2000) – typically assume that there will be some gradual convergence of income levels towards those of most developed economies. A similar approach is taken here, but special emphasis is put on integrating some of the current theoretical and empirical knowledge on long-term economic growth, and making transparent assumptions about the drivers of GDP growth over the projection period (for discussion of assumptions, detailed results and data sources, see Duval and de la Maissonneuve, 2010).

A “conditional convergence” hypothesis is incorporated into the projections. Following past research (*e.g.* Hall and Jones, 1999; Easterly and Levine, 2001), and based on a standard aggregate Cobb-Douglas production function with physical capital, human capital, labour and labour-augmenting technological progress, GDP per capita is first decomposed as follows for 2005:

$$Y_t / Pop_t = (K_t / Y_t)^{\alpha/(1-\alpha)} A_t h_t (L_t / Pop_t)$$

where Y_t/Pop_t , K_t/Y_t , A_t , h_t , and L_t/Pop_t denote the level of GDP per capita (using PPP exchange rates to convert national GDPs into a common currency), the capital/output ratio, total factor productivity (TFP), human capital per worker and the employment rate, respectively. α is the capital share in aggregate output.

Based on this, long-term projections are then made for each of the four components so as to project the future path of GDP per capita:

- Long-term annual TFP growth at the “frontier”, defined as the average of the “high-TFP” OECD countries, is 1.5%. The speed at which other countries converge to that frontier is assumed to tend gradually towards 2% annually.
- Where it is currently highest, the human capital of the 25-29 age group is assumed to level off, based on past experience. The speed at which other countries converge to that frontier is assumed to tend gradually towards a world average between 1960 and 2000. The human capital of the working-age population is then projected by cohorts.
- Capital/output ratios in all countries gradually converge to current levels in the United States, which is implicitly assumed to be on a balanced growth path. In other words, marginal returns to capital converge across countries over the very long term in a world where international capital is mobile.
- Employment projections combine population, participation and unemployment scenarios. The United Nations population projections are used (baseline scenario). In those OECD countries where participation is currently highest, future retirement ages are partially indexed to life expectancy. Elsewhere, participation rates gradually converge to the average in “frontier” countries. Unemployment rates converge to 5%.

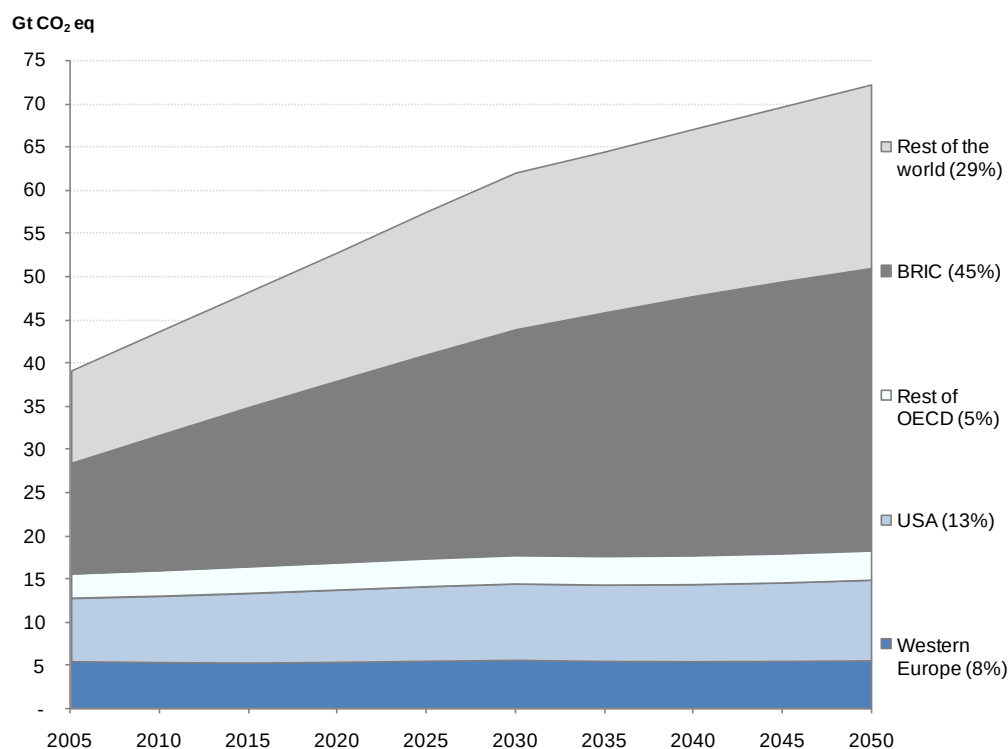
This framework was applied to 76 countries, covering 90% of the world’s GDP and population in 2005. For all other countries, the productivity convergence scenario to labour productivity or GDP per capita was applied instead of TFP. The approach followed addresses recent criticisms of economic

projections using market exchange rates, which form the vast majority of scenarios in the literature (Castles and Henderson, 2003a, 2003b; Henderson, 2005). This is achieved in two ways: (i) By using purchasing power parities (PPPs), not market exchange rates, to compare initial income per capita levels; (ii) by assuming faster future productivity growth in tradable than in non-tradable industries, in line with historical patterns. Reflecting this “Baumol-Balassa-Samuelson” effect, the real exchange rate of fast-growing countries typically appreciates. Therefore, the GDP PPP per worker path produced by the ENV-Linkages model combines both a volume effect (GDP growth in constant national currency) and a relative price effect (the real exchange rate appreciation), with the former being the main driver of emissions.

Assumptions about other drivers of emissions

The BAU scenario was developed on the basis of the pre-crisis surge of the international crude oil price, and therefore assumed that it would culminate at USD 100 per barrel (in real 2007 prices) in 2008, stay constant in real terms up to 2020 and increase steadily thereafter up to USD 122 per barrel in 2030. Beyond that horizon, oil exporters’ crude oil supply is projected to decelerate gradually, roughly reflecting reserve constraints, and resulting in a sustained rise in the real crude oil price beyond 2030 at 1% annually between 2030 and 2050. The international price of natural gas is assumed to follow the international crude oil price up to 2030, but this link then weakens somewhat, reflecting a higher assumed long-term supply elasticity for natural gas than for oil. Coal prices are projected to rise only modestly (in real terms) beyond their recent levels. The price of steam coal is assumed to reach USD 100 per tonne in 2008, in line with the assumption of a high long-term supply elasticity. International Energy Agency (IEA, 2008) energy demand projections were used to calibrate future energy efficiency gains. These assume a gradual weakening of the relationship between economic growth and energy demand growth, especially after 2030.

Figure A2.1. Projected GHG emissions¹ by country/region (2005-50, Gt CO₂eq)



Note: Countries/regions in this figure are based on the 12-regions aggregation of the ENV-Linkages model. Korea, Mexico and Turkey are included in the Rest of the World (ROW).

1. Excluding emissions from Land Use, Land-Use Change and Forestry. Number in brackets represents percentage share of total emissions in 2050.

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